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Toxicological studies of engineered nanoparticles in biological systems

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A B S T R A C T

Recently the ability of engineering of nanoparticles and to produce nanomaterials at the nano or near-atomic scale has triggered their rapid production due to their interesting properties that were not previously seen at scales above the micrometers. The exponential demands of nanosized materials with use of nanoparticles have been increased in industrial applications. With increasing use of nanoparticles in a variety of consumer goods, biological systems are constantly exposed to such nanomaterials besides exposure at production sites. This unintended exposure to nanomaterials may occur *via* inhalation, dermal exposure or gastrointestinal tract absorption and may pose a great risk to environment. However, there are still widespread controversies and ambiguities with respect to the toxic effects and mechanism of action of engineered nanoparticles in biological systems. The present study reviewed toxicological studies of engineered nanoparticles in biological systems.

Introduction

Nanoparticles (NPs) or Nanomaterials are the materials having the size less than 100nm that can be nanoscale in one dimension (e.g. films), two dimensions (fibres and tubes) or three dimensions (particles). Many theories have been proposed to explain the toxicity phenomenon of nanoparticles including, generation of Reactive Oxygen Species (ROS), which can disrupt cell structures, interference with normal metabolism (Nel *et*

al. 2006, Choi and Hu, 2009), binding with macromolecules making them dysfunctional (Gogoi *et al.* 2006). However, the real impact of nanoparticles and their mechanism of reactions against biological system are still not known. Researchers have provided evidence for NP-mediated production of ROS and generation of oxidative stress as a possible mechanism of toxicity (Oberdorster, 2005, Pickering and Wiesner, 2005, Zeng *et al.* 2015) especially for

carbonaceous nanoparticles (*i.e.* fullerenes, fullerols and carbon nanotubes) and nanoparticulate like titanium dioxide and zinc oxide (Reeves *et al.* 2008). NPs of TiO₂ and ZnO are widely used in sun care products (Serpone *et al.* 2007), as well as on self cleaning coatings (Cai *et al.* 2008) that may be released into the environment and become a threat to ecosystems.

Toxicity of NPs in Animal Model System

Federici *et al.* (2007) observed series of physiological effects induced by TiO₂-NPs in Rainbow trout *Oncorhynchus mykiss*. Further, Baun *et al.* (2008) observed toxic effects of TiO₂ and ZnO-NPs on algae *Pseudokirchneriella subcapitata*, Crustaceans *Daphnia magna* and *Thamnocephalus platyurus* and bacteria *Vibrio fischeri* (Heinlaan *et al.* 2008). Toxicity of nanoparticles of silver, platinum, carbon nanotubes on terrestrial animals was also observed by Dani *et al.* (2008); Johansen (2008); Petersen *et al.* (2009) and Reeves *et al.* (2008). Invertebrates are able to intake nanomaterials dispersed in the environment by different ways: *direct ingestion; from contaminated preys; water filtration; inhalation* and *surface contact*. Some degree of bio-modification was also observed in daphnids (Oberdorster *et al.* (2005); Roberts *et al.* (2007) and Baun *et al.* (2008). Compartmentalization of nanosized contaminants in selected tissues and in organelles has been documented by Moore *et al.* (2006); Li *et al.* (2011); Jaiswal *et al.* (2003); Kohler *et al.* (2008) and Oberdorster *et al.* (2002).

The Rotifer *Brachionus Calyciflorus* was studied under exposure to nanomaterials as a component of simplified food web, algae and bacteria. Those were predated by a protozoans which in turn was the rotifer

prey and both were fed by fish (Holbrook *et al.* 2008). Acute and chronic toxicity was observed by Borgmann *et al.* (2005) in *Hyalella azteca* (Family Hyalellidae) and *Leptocheirus plumulosus* (Family Aoridae) shrimps inhabited on fresh water and feed on sediment or suspended particles. Griffith *et al.* (2008), reported lethality in three different aquatic organisms namely, algae (*Pseudokirchneriella subcapitata*), Daphnids (*Daphnia pulex*, *Ceriodaphnia dubia*) and fish (*Danio rerio*) on exposure of Ag-NP and Cu-NP. They also demonstrated that copper and silver nanoparticles are acutely toxic to wide spectrum of aquatic species including zebra fish. This toxicity is largely manifest at the gills and were not affected simply by particles dissolution. Moore (2006) stated that engineered nanoparticles are likely to be deposited in aquatic systems and represent a possible danger to aquatic life. Both dissolved and particulate nanoparticles enhanced deposition of metals in gills. Branchial uptake of ionic silver and copper nanoparticles has been well documented in freshwater fish by Bury and Wood (1999) and proposed several mechanisms to enhanced silver levels in gills. NPs may be trapped in the mucus layer of the gill as demonstrated for larger particles by Dani *et al.* (2008) and Connor *et al.* (2005).

The sediment particles less than 500 nm were observed intracellularly in salmonid gill epithelial cells by Martens and Seruizi, (1993). Morgan *et al.* (2004) observed significant increase in whole body silver content on exposure of silver NPs. However, it is not clear either is due to translocation of silver through gills or due to ingestion of particulates and gastro Intestinal (GI) absorption. Soluble copper is a well known gill toxicant in freshwater fish species (Mazon *et al.*, 2002) and causes significant thickening in gill filament (Griffith *et al.*

2008). Morgan *et al.* (1997) reported silver as most potent gill toxicants in freshwater fish and caused highly specific inhibition of Na⁺/K⁺ ATPase. Dabbousi *et al.* (1997) observed change in morphology in spiny dogfish, a marine elasmobranch on exposure of Ag-NPs.

Common Animal Model System used for Nanotoxicity Test

Usually Daphnids (*Daphnia magna*, *Daphnia pulex* and *Ceriodaphnia dubia*) preferred as first choice for ecotoxicological tests of nanoparticles (Joncxyk and Gilron, 2005). Benthic organisms (cnidarian *Hydra attenuate*) are also used (Dabbousi *et al.*, 1997; Blaise and Kusui 1997; Holdway, 2005) for long term toxicity studies. Their clubbing movements of tentacles are observed as early signs of exposure to toxic agents, before reproductive changes or death.

Impact of Metal Nanoparticles on Earthworms

Mc Shane *et al.* (2010) recorded the exposure of 10,000mg TiO₂/kg soil not significantly affected the biomass of *Eisenia spp.* In contrast, Heckmann *et al.* (2008) reported a 49% reduction in the number of juveniles produced by *E. fetida* following 28d of exposure to 1,000 mg TiO₂/kg soil. Recent nano-ecotoxicological studies indicates so far that increasing dose provides an increase in response, although this may not follow traditional mass based dose-response relationships (Oberdorster *et al.* 2005). Ni-NP had a much lower toxicity to *E. fetida* than NiCl₂ possibly due to a strong agglomeration of the Ni particles combined with little or no oxidation of the NP-Ni particle (Besenbacher and Norskov, 1993). There was also a tendency of increased biomass was reported by Gurr *et al.* (2005)

for Cu-NP, and Ni-NP exposed worms. Very limited information is available for toxicity tests of nanoparticles (TiO₂-NP; Ag NO₃, Al₂O₃, SiO₂, TiO₂, ZnO₂) against earthworms except few including Hu *et al.* (2010); Petersen *et al.* (2008). Oberdorster *et al.* (2004) demonstrated NPs may affect soil ecosystem *via*: 1) direct effect; 2) changes in the bioavailability of toxins or nutrients; 3) indirect effects resulting from their interaction with natural organic compounds; 4) interaction with toxic organic compounds which may amplify or alleviate their toxicity. It was also observed earthworms are able to differentiate particles through unknown mode of action (McShane *et al.*, 2006). Gomes *et al.* (2012) demonstrated NPs altered mRNA levels for specific genes mainly involved in metabolism, transcription and translation or in the stress response and generates oxidative stress conditions. The reduction in growth and development, damaged cuticle with underlying pathologies of epidermis, muscles and gut barrier on exposure of fullerene NPs (C60) to *Lumbricus rubellus* was reported by Vander Ploeg *et al.*, (2013). Li *et al.* (2011) observed *Eisenia fetida* shown neither response of antioxidant enzyme expression or activity nor acute toxicity in C60 spiked soil. We studied effect of ZnO-NPs (Gupta *et al.*, 2014). on *E. fetida* in terms of reproductive behavior, antioxidant enzyme activities and accumulation of Zn⁺⁺ remote from portal of entry. It may concluded that interaction of nanoparticles with earthworm is unpredictable that may result in ecologically significant effects on behavior at environmentally relevant concentrations.

Biological Uptake and Mechanisms of toxicity

Published quantitative research on uptake and accumulation of NPs by whole

organisms is very limited. It is known that organisms living in NPs containing environments incorporate them within their bodies, mainly *via* the gut (Baun *et al.* 2008; Roberts *et al.* 2007) with a possibility of translocation within the body. Most of the initial work, in this area was undertaken on standard animal models (daphnids) used in ecotoxicology. Feng *et al.* (2005) demonstrated the uptake of fluorescent carboxylated NPs by *Daphnia magna* and translocation from the gut to reserve fat droplets. Some NPs such as quantum dots and CNTs (carbon nanotubes) are intentionally designed to interact with proteins, nucleic acids or cell membranes for labeling or drug delivery purposes (Gao *et al.* 2008; Medintz *et al.* 2008). The toxicity mechanisms have not yet been completely elucidated for nanoparticles. Possible mechanisms includes disruption of membranes or membrane potential, oxidation of proteins, genotoxicity, and interruption of energy transduction, formation of ROS and release of toxic constituents.. Meng *et al* (2007) reported NPs first, adhered to the surface alter the membrane properties, therefore affecting the permeability and the respiration of the cell, they can penetrate inside cell and caused DNA damage and they can also release toxic Ag⁺ ions. Degradation of lipo-polysaccharide molecules forms pits in the membrane that changes membrane permeability due to silver nanoparticles as reported by Moore *et al.* (2006). Gold nanoparticles (AuNP) can easily enter into cells (Connor *et al.* 2005) and bind strongly to amine and thiol groups of cells that enabled their surface modification with amino acids and proteins (Oberdorster 2001; Ville *et al.*, 1995).

Impact of NPs in biological system

Damage to membrane integrity

The imaging applications of NPs are mainly targeted on cell membrane. NPs attached to

cell surface and compromise with the integrity and functions of the cell membrane. Silicon (Si) NPs and fullerene derivatives can embed themselves in the membranes (Zang *et al.* 2015). Carboxyfullerene puncture bacterial cell membrane in a gram positive bacterial strain and causes cell death (Huang *et al.*, 2001). Ingle *et al.* (2008) reported Au-NPs weaken cell membranes and causes heat shock responses in *Escherichia coli*. NPs can also indirectly causes membrane damage through the generation of ROS which can oxidize double bonds on fatty acid tails of membrane phospholipids in a process known as lipid peroxidation. This increases membrane permeability and fluidity, making cell more susceptible to osmotic stress or hindering nutrient uptake (Cabisco *et al.*, 2000). Peroxidized fatty acids may trigger reactions that generate other free radicals leading to more cell membrane and DNA damage (Valembois *et al.*, 1994).

Protein destabilization and oxidation

NPs-protein interactions have been optimized for a variety of biomedical applications. As quantum dots are used to target and fluorescently label proteins for imaging (Jaiswal *et al.* 2003; Medintz *et al.* 2005). Oberdorster *et al.* (2004) observed ROS can also lead to the formation of disulfide bonds between sulfur containing amino acids thus disturbing the structure and function of the protein.

Nucleic Acid damage

Interactions of NPs with nucleic acids have applications in DNA labeling or DNA cleavage. The tagged nucleotides with NPs especially quantum dots commonly used as labeling agents for bioimaging applications (Dyadyusha *et al.*, 2005; Dubertret *et al.*, 2002). Cabisco *et al.* (2000) observed some

NPs indirectly damage DNA because of ROS production, which breaks DNA strand, cross linking and adducts of bases or sugars. Mson *et al.* (2006) stated TiO₂NPs generates oxygen radicals that can nick supercoiled DNA. Scientific committee on Emerging and Newly Identified Health Risks, (2007) shown possibility of tumor formation through DNA damage and increase in cell proliferation associated with inflammation.

Cell disruption due to ROS

NPs induces ROS that may damage every cell component, and this interaction tends to trigger further radical formation.

Interruption of Energy transfer

Electron transport phosphorylation and energy transduction processes may be disrupted if membrane integrity is compromised or if a sensitive NPs contacts membrane bound electron carriers and withdraws electrons from the transport chain. Fullerene derivatives have been reported to inhibit *E. coli* respiration of glucose by Mc Shane and Ritter (2010).

Release of toxic components from NPs

Certain NPs causes toxicity to bacterial cells by releasing harmful components, such as heavy metals or ions. Quantum dots are semiconductor nano-crystals that contain noble or transition metals such as CdSe, CdTe, CdSeTe, Zn Se. In As or PbSe in their Core, CdS or ZnS in their shell, and an organic coating (Dubertret 2002). Release of silver ion has been implicated in toxicity of silver NPs. It is believed that silver ions interact with thiol groups of proteins, resulting in inactivation of vital enzymes (Matsumura *et al.*, 1996). Silver ions have also been shown to prevent DNA replication

and affect the structure and permeability of the cell (Feng *et al.* 2000).

Conclusions

We discussed the toxicity studies of engineered nanoparticles in biological systems. Researchers observed toxicity of nanoparticles at different levels and it depends especially, on size and composition of nanoparticles. However, the mechanism of their action in biological system is not very clear. It is essential to decrease toxicity of engineered nanoparticles at non-significant level to ensure safe application of nanoparticles to organisms and to the environment. Further, understanding of human health implications and ecological consequences in environment at exposure of unintended engineered nanoparticles is essential before the commercial benefits of these nanomaterials fully realized.

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